

**ifs**

# International Fertiliser Society

## **EFFECTS OF UREASE AND NITRIFICATION INHIBITORS ON YIELDS AND EMISSIONS IN GRASSLAND AND SPRING BARLEY**

by

**Patrick J. Forrester<sup>1</sup>, David Wall<sup>1</sup>, Rachael Carolan<sup>2</sup>, Mary Harty<sup>1</sup>,  
Leanne Roche<sup>1</sup>, Dominika Krol<sup>1</sup>, Catherine Watson<sup>2</sup>, Gary Lanigan<sup>1</sup>  
and Karl Richards<sup>1</sup>**

<sup>1</sup>Teagasc, Soils and Land-Use Department, Crops, Environment and  
Land Use Programme, Johnstown Castle, Co. Wexford, Ireland.

<sup>2</sup>Agri-Food and Biosciences Institute (AFBI), Newforge Lane,  
BT9 5PX, Belfast, United Kingdom.

**Proceedings 7xx**

Paper presented to the International Fertiliser Society  
at a Conference in Cambridge, United Kingdom, on 9<sup>th</sup> December 2016.

[www.fertiliser-society.org](http://www.fertiliser-society.org)

© 2016 International Fertiliser Society

ISBN 978-0-85310-3nn-n

(ISSN 1466-1314)

## SUMMARY.

In trials conducted in the temperate maritime climate of Ireland on a range of acidic soils, calcium ammonium nitrate (CAN) and urea gave comparable yield performance. There was little evidence of reduced yields by using urea for grassland or spring barley. Our finding that urea produced annual yields that were not significantly different from CAN differs from previous studies which found that yields from urea were lower than those from ammonium nitrate or nitrate based fertiliser in the UK. However, there are also published results from trials conducted in temperate Irish grassland showing equal yield performance of CAN and urea in the 1970s. Based on yield performance and the cost of fertiliser there is scope to dramatically increase the level of urea usage in straight and blended fertilisers in the temperate maritime climate of Ireland in both grassland and spring barley. Such an increase will bring substantial benefits in terms of reducing direct nitrous oxide ( $N_2O$ ) emissions from fertiliser applied to soil, particularly in poorly draining soils subject to high levels of precipitation. Nitrogen recovery by plants tends to be more sensitive to differences in fertiliser efficiency than is yield. Although yields did not differ between urea and CAN; urea had a lower nitrogen recovery indicating that urea usage will also result in a reduced level of fertiliser use efficiency. Reduced efficiency is less tangible to farmers who tend to be primarily concerned with dependable yield results. Reduced efficiency is a problem nonetheless, particularly as it is closely linked to  $NH_3$  emissions in urea usage. European countries including Ireland have committed to reduce national  $NH_3$  emissions to comply with the revised National Emission Ceilings Directive (2001/81/EC) in Europe. Increased urea usage, which looks attractive from a yield, cost and direct  $N_2O$  perspective in Ireland, runs counter to meeting these commitments. Additionally,  $NH_3$  is a source of indirect  $N_2O$  emissions that will negate some of the  $N_2O$  savings from urea. Due to the issues of yield dependability, fertiliser efficiency,  $N_2O$  and  $NH_3$  emissions the urease inhibitor N-(n-butyl) thiophosphoric triamide (NBPT) is a particularly attractive option for making urea use more efficient by addressing its key weakness in the area of variable  $NH_3$  loss and efficiency. The urease inhibitor NBPT along with the nitrification inhibitor dicyandiamide (DCD) were tested with urea in comparison with calcium ammonium nitrate (CAN). The nitrification inhibitor DCD was very effective in reducing fertiliser N associated  $N_2O$  emissions. Indeed, its usage allowed  $N_2O$  levels to be reduced to levels comparable to where no application of N fertiliser was made at some site-years. However, at the DCD incorporation rate tested, DCD contributed to variability in  $NH_3$  loss from urea and suppressed both yield response and fertiliser efficiency. Use of the urease inhibitor NBPT in addition to DCD went a substantial way to resolving these shortcomings. Continuing work is needed to tailor the rate of existing and new urease and nitrification inhibitors to optimise the balance between suppression of gaseous N emissions, agronomic performance and economic considerations.

## CONTENTS

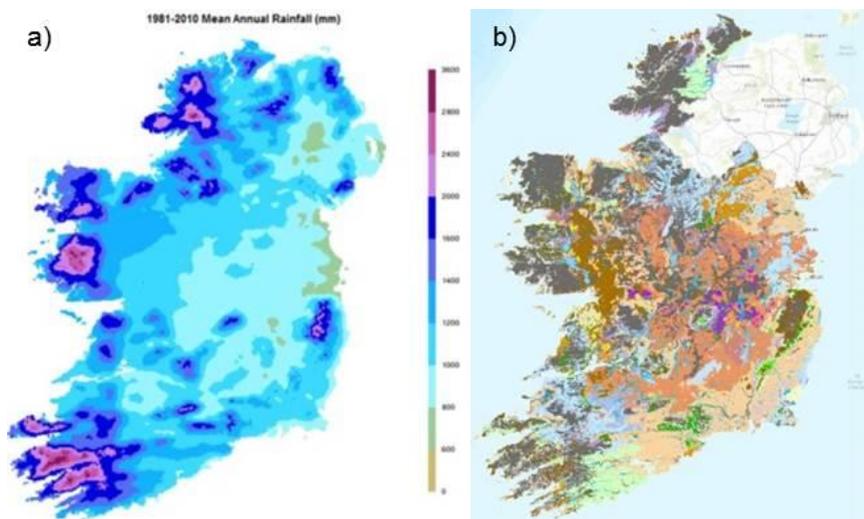
Summary	2
1. Introduction	4
2. Materials and methods	5
2.1 Grassland sites	5
2.2 Spring barley site	7
3. Results and discussion	8
3.1 Grassland	8
3.1.1 Grassland yields	8
3.1.2 Nitrogen uptake and apparent fertiliser recovery in grassland	8
3.1.3 Ammonia emissions from grassland	11
3.1.4 Nitrous oxide emissions from grassland	12
3.2. Spring barley	13
3.2.1 Spring barley yields	13
3.2.2 Spring barley apparent fertiliser nitrogen recovery	14
3.2.3 Nitrous oxide emissions from spring barley	15
4. Conclusions	16
4.1 Yield	16
4.2 Efficiency: apparent fertiliser nitrogen recovery (AFR)	16
4.3 Ammonia	17
4.4 Nitrous oxide	17
5. Acknowledgements	17
6. References	18
Related Proceedings of the Society	21

*Agrotain<sup>®</sup> is a registered trade mark of Koch Agronomic Services, Wichita, KS, USA.*

**Keywords:** Fertiliser nitrogen, urease inhibitor, nitrification inhibitor, nitrous oxide, ammonia, grassland, spring barley, yield, efficiency.

## 1. INTRODUCTION.

Application of nitrogen (N) fertiliser is a cornerstone of many intensive agricultural systems, including those in Ireland, and is a key input moving crops closer to fulfilment of their genetic yield potential. It is estimated that N and other mineral fertilisers feed around half of the global population (Sutton *et al.*, 2013). A key challenge is how to continue to apply fertiliser N to underpin crop yields while curtailing reactive N losses, including gaseous N emissions from fertiliser applied to diverse soils under varying weather conditions. Varying soils and precipitation patterns (Figure. 1) increase this challenge in environments such as Ireland. The challenge of gaseous N loss has come into particular focus recently due to national commitments across Europe to reduce losses of  $\text{NH}_3$  and greenhouse gases (GHGs), including  $\text{N}_2\text{O}$ . Nitrous oxide comprises approximately 32% of agricultural GHG emissions (U.S. EPA, 2012) and is a potent GHG, with a global warming potential 265 times greater than  $\text{CO}_2$  over a 100 year time frame and  $\text{N}_2\text{O}$  has an atmospheric lifetime of 121 years (IPCC, 2014).



**Figure 1:** a) Mean annual precipitation (mm) (source Met Eireann) b) the soils of Republic of Ireland (source: the Irish Soil Information System).

The use of urea in place of ammonium nitrate-based fertiliser in intensive grassland has been linked to reduced direct  $\text{N}_2\text{O}$  emissions in cool wet soils (Dobbie and Smith, 2003). However, urea is vulnerable to  $\text{NH}_3$  volatilisation (Chambers and Dampney 2009; Forrestral *et al.* 2016). Ammonia volatilisation is problematic from the perspective of reducing national  $\text{NH}_3$  losses and also represents a source of indirect  $\text{N}_2\text{O}$  emissions as  $\text{NH}_3$  is re-deposited and nitrified. A urease inhibitor can reduce  $\text{NH}_3$  volatilisation from urea by inhibiting the enzyme urease which catalyses urea hydrolysis. The urease inhibitor NBPT has been shown to reduce  $\text{NH}_3$  loss from urea (Watson *et al.*, 1990, 1994; Goos *et al.*, 2013). Working on a different pathway, nitrification inhibitors, such as DCD, inhibit ammonia mono-oxygenase, which catalyses oxidation of ammonium

( $\text{NH}_4^+$ ) to nitrite (Kim *et al.* 2012). Although effective for reducing  $\text{N}_2\text{O}$  emissions and leaching, nitrification inhibitors may increase  $\text{NH}_3$  emissions. In a meta-analysis, Kim *et al.* (2012) reported that the effect of nitrification inhibitors on  $\text{NH}_3$  emissions was inconsistent, with increased  $\text{NH}_3$  emissions in 26 studies, no change in 14 studies and decreased emissions in six studies. Lam *et al.* (2016) suggest that the beneficial effects of nitrification inhibitors in decreasing direct  $\text{N}_2\text{O}$  emissions may be undermined or outweighed by increased  $\text{NH}_3$  emission and the associated indirect  $\text{N}_2\text{O}$  emission.

This paper will examine the results of recent trials assessing the effects of fertiliser type, urease and nitrification inhibitors on yields, fertiliser efficiency and gaseous emissions in the temperate maritime climate of Ireland. In the context of sustainable intensification of Irish agriculture, where environmental protection and economic competitiveness are equal and complementary (FoodWise 2025 in the Republic of Ireland and Going for Growth 2020 in Northern Ireland), solutions for the challenge of using fertiliser N with reduced reactive N losses including gaseous losses are particularly in focus. In Ireland, agriculture must take action for national  $\text{NH}_3$  and GHG reduction commitments to be realised. In the Republic of Ireland agriculture accounts for c.98% of national  $\text{NH}_3$  emission and c.32% of GHG emissions. Northern Ireland has a similar agricultural emissions profile with agriculture accounting for c.93% of  $\text{NH}_3$  emissions and c.29% of GHG emissions. Fertiliser N application is associated with emissions of  $\text{NH}_3$  and  $\text{N}_2\text{O}$ , making the use of N sources and/or inhibitors capable of reducing these reactive N losses particularly relevant.

Currently, the main straight fertiliser N used in Ireland is CAN with urea and urea amended with a urease inhibitor taking up a smaller portion of the market.

## **2. MATERIALS AND METHODS.**

### **2.1. Grassland sites.**

To measure the effect of fertiliser N type on grass yield,  $\text{N}_2\text{O}$ ,  $\text{NH}_3$  and soil mineral N, replicated field experiments were conducted at three grassland sites in Ireland (Figure 2) in 2013 and 2014 (six site-years). The locations were Johnstown Castle, Co. Wexford, Moorepark, Co. Cork and Hillsborough, Co. Down. The site details are provided in Table 1. The sites were chosen to represent a range of soil and geo-climatic conditions across intensively managed agricultural areas in Ireland.

The experimental design was a randomised complete block with five replicates at each site-year. The CAN, urea and urea+NBPT fertiliser treatments were applied at annual N rates of 100, 200, 300, 400 and 500 kg/ha in five equal split applications between March and September. Plots received a basal application of P, K, and S in line with soil test recommendations to ensure that these nutrients were not limiting. Soil pH levels at the experimental sites were below the recommended level of 6.3. However, no lime was applied to the experimental sites to avoid the confounding effects of liming on the performance of the urea fertiliser (Watson *et al.*, 1987). Of the 28,137 grassland soil samples received for testing by Teagasc in 2015 64% were below pH 6.3. Indeed, 46% of grassland soils received by Teagasc were below pH 5.9.

**Table 1:** *Climate and soil physical and chemical properties for the six site-years tested (adapted from Harty et al., 2016).*

Site-year	HB, 2013	HB, 2014	MP, 2013	MP, 2014	JC, 2013	JC, 2014
GPS coordinates	54°27'827N, 6°04'57873W	54°45'127N, 6°04'5785W	52°9'27"N, 8°14'42"W	52°9'33"N, 8°14'43"W	52°18'27N, 6°30'14W	52°17'32"N, 6°30'7"W
Drainage <sup>a</sup>	Imperfect	Imperfect	Good	Good	Good	Moderate
Soil pH	5.7	5.6	5.6	5.4	5.5	5.7
Soil texture	Clay loam	Clay loam	Sandy loam	Sandy loam	Sandy loam	Sandy loam
Sand (%)	41.0	42.9	58.8	57.8	52.9	51.7
Silt (%)	33.9	34.6	27.8	28.2	33.2	33.9
Clay (%)	25.1	22.4	13.5	14.1	13.9	14.4
BD	0.86 <sup>1</sup>	0.79	1.02	1.18	1.11	1.27
Soil TC (%)	5.99	5.16	3.00	3.02	3.16	2.83
TN (%)	0.557	0.451	0.318	0.321	0.304	0.284
Soil LOI (%)	14.3	12.5	7.40	7.90	7.30	7.02
CEC (meq/100g)	28.5	25.4	16.7	18.4	15.6	15.5
Rainfall (mm)						
Annual	1,113	1,047	1,130	1,002	1,021	939
30-year average	885	885	1,018	1026	1,060	1,060
Main growing season <sup>b</sup>	560	478	407	459	336	441
30-year average growing season	478	478	509	512	534	534

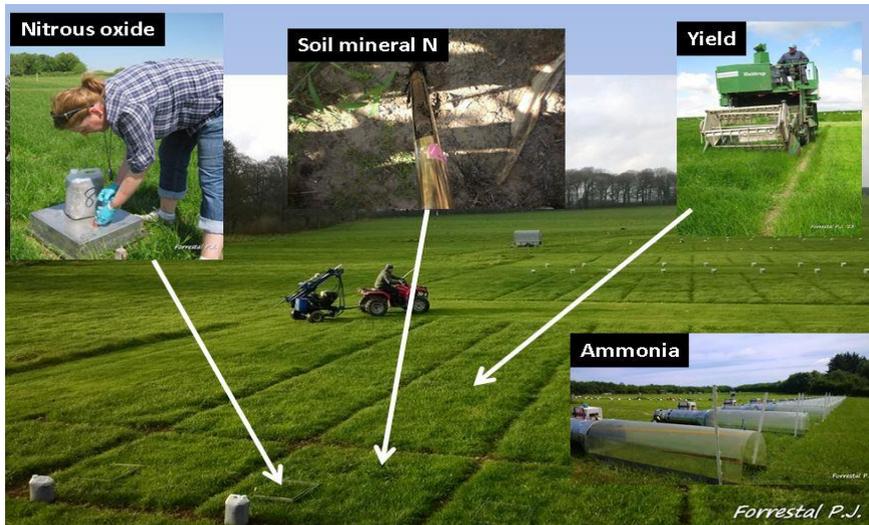
<sup>a</sup> Drainage Classification was based on the soil associations from the Soil Map of Ireland (Gardiner and Radford, 1980).

<sup>b</sup> Main growing season (1 March to 30 Sept).

Urea+DCD and urea+NBPT+DCD were applied at the 200 kg N/ha rate only. In addition there was a zero N control treatment. The source of the urease inhibitor NBPT was Agrotain<sup>®</sup>, which was coated onto urea granules at 660 ppm NBPT (on a urea weight basis). Watson *et al.* (2008) reported little additional benefit from exceeding 250 ppm. However, based on their work, the 660 ppm NBPT level was expected to achieve near maximum NH<sub>3</sub> loss abatement. For the urea+DCD treatment the DCD had been added to urea at the molten stage at the rate of 3.5% on a urea-N basis. Urea+DCD granules were coated with Agrotain<sup>®</sup> on site to a rate of 660 ppm NBPT to make urea+NBPT+DCD.

Nitrous oxide emissions were measured from each plot using the static chamber technique (Chadwick *et al.*, 2014). Yield and N uptake was measured by harvesting dedicated agronomic plots (2 m x 10-12 m at the end of each grass growth cycle (Figure 2). Soil mineral N was measured by sampling the dedicated soil sampling area of the plots. Ammonia emissions were measured from each of the fertiliser treatments at the Johnstown Castle and Hillsborough sites during 2014 using a system of wind tunnels (Lockyer, 1984; Meisinger *et al.*, 2001). Each wind tunnel covered an area of 0.5 x 2 m. There were three replications per treatment. Ammonia measurements were

conducted adjacent to those used for N<sub>2</sub>O and yield measurements with fertiliser applications applied at the same timing as for the other plots (Figure 2). For the detailed experimental setup see Forrestral *et al.* (2016) and Harty *et al.* (2016).



**Figure 2:** The arrangement of each experimental unit in the grassland site.

## 2.2. Spring barley site.

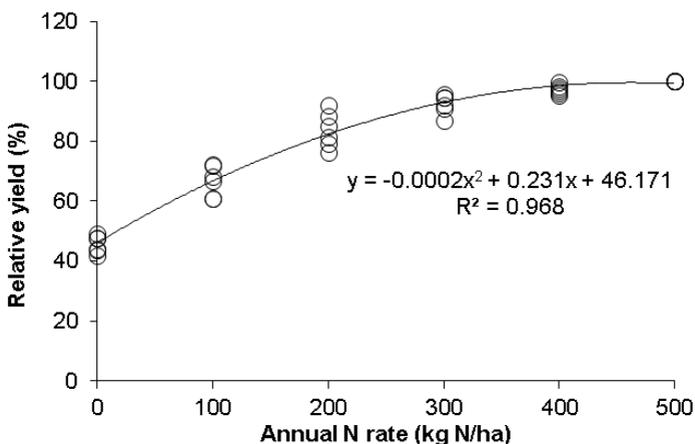
The spring barley trial was conducted on one free-draining loam site located in Marshalstown, Co. Wexford (52° 33' 37.3" N 6° 36' 09.0" W) over three years. The experimental design was a randomised complete block with five replicates. The layout of each experimental unit was similar to the grassland sites (Figure 2) with a dedicated N<sub>2</sub>O, soil mineral N and yield measurement area. The same suite of fertilisers used in the grassland experiment were applied in two split applications. The first split application of 30 kg N/ha was surface applied shortly following planting. The second and final split application to balance to the target N rate was applied at early to mid-tillering. Nitrous oxide emissions were measured using the static chamber technique (Chadwick *et al.*, 2014). The soil pH was 6.8, total C 2.9% and the CEC 100 meq/100g (Roche *et al.*, 2016). Forty five percent of tillage soils received for testing by Teagasc in 2015 had a pH >6.5. The field site where experiments were conducted was located within a major malting barley growing region in Ireland and was representative of the typical soil type used for spring barley production in Ireland. The site had been in long-term arable production for the previous 20 years.

### 3. RESULTS AND DISCUSSION.

#### 3.1. Grassland.

##### 3.1.1. Grassland yields.

Figure 3 summarises the response to N fertiliser rate for the six site-years tested under a cutting regime and demonstrates that grassland responds strongly to N addition (Figure 3). For the three fertiliser types tested over a range of N rates (CAN, urea and urea+NBPT) there was no significant interaction between N rate and fertiliser type nor was there a main effect of fertiliser type on yield (Forrestal *et al.*, in review). Across this range of rates the three fertilisers produced yield which did not differ significantly. Our finding that urea produced annual yields that were not significantly different from CAN differs from previous studies which found that yields from urea were lower than those from ammonium nitrate or nitrate based fertiliser in the UK (Rodgers *et al.*, 1984; Chaney and Paulson, 1988). However, our results are not unprecedented in Irish temperate grassland. Keane *et al.* (1974) also found that urea yielded as well as CAN in Irish temperate grassland in the 1970s.

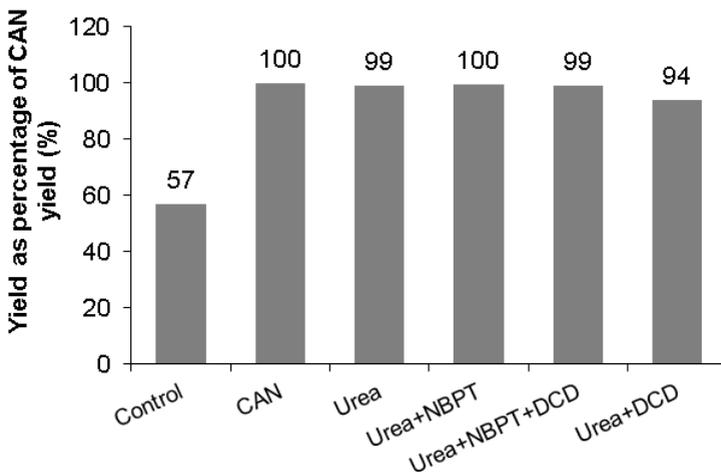


**Figure 3:** Grassland yield relative to the 500 kg N/ha rate for six site-years under a cutting regime. No significant N fertiliser type effect was detected at  $P < 0.05$ . Adapted from Forrestal *et al.* (in review).

A larger suite of fertilisers was examined at the 200 kg N/ha/yr rate. The 200 kg N/ha rate is at a point in the response curve where these grassland sites were still responding strongly to N fertiliser addition under a cutting regime (Figure 3). Grass yield was similar for all fertiliser treatments except for urea with the nitrification inhibitor DCD alone (Figure 4). The meta-analysis of Kim *et al.* (2012) indicates that nitrification inhibitors such as DCD may increase  $\text{NH}_3$  loss from urea. In this study, when NBPT was added to DCD treated urea, grass yields were similar to the highest yielding treatments. This result lends weight to the hypothesis that the  $\text{NH}_3$  loss pathway is an important reason why urea+DCD alone had poorer yield performance than the other fertiliser treatments.

### **3.1.2. Nitrogen uptake and apparent fertiliser recovery in grassland.**

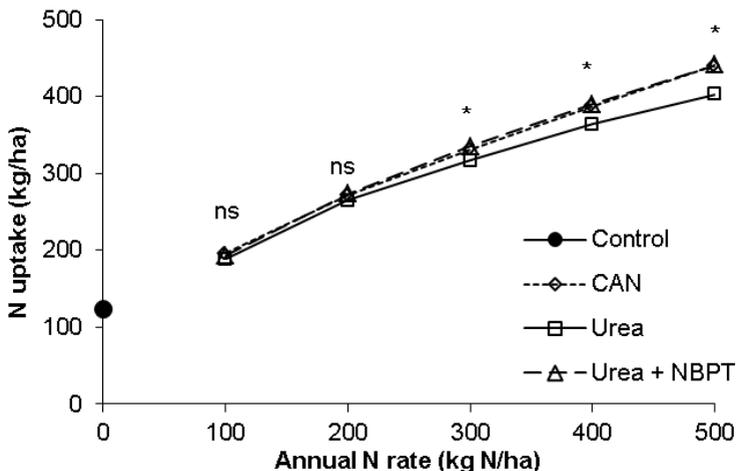
Urea is vulnerable to  $\text{NH}_3$  loss in temperate grassland (Watson *et al.*, 1990) and the proportion of  $\text{NH}_3$  loss tends to increase with increasing urea rate (Chambers and Dampney, 2009; Forrester *et al.*, 2016). Despite this, in these trials, no significant yield difference was detected between urea and CAN or



**Figure 4:** Average grassland yield expressed as a percentage of CAN yield. Data from six site-years fertilised at 200 kg N/ha adapted from Harty *et al.* (in press).

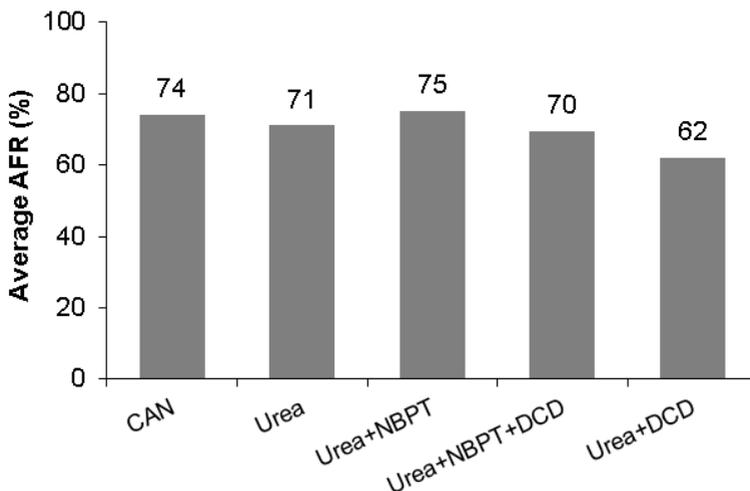
urea+NBPT even at high N rates (Figure 3). Nitrogen uptake and recovery by plants is more sensitive to detection of differences in N availability between treatments than is yield. Figure 5 demonstrates this point. Significantly lower N uptake was measured for urea compared with both CAN and urea+NBPT at higher N rates whereas no significant yield difference was detected at these N rates (Figure 3). It is worth noting that all N rates were applied in five split applications ranging from 20 to 100 kg N/ha/split. In a silage setting the higher N rates are likely to be applied by a farmer at least once in a growing season.

Comparison of the full suite of fertiliser products at the 200 kg N/ha/yr rate also provides insight into the efficiency of the various fertiliser products. Across the six site-years urea+NBPT and CAN had the highest apparent fertiliser N recovery (AFR). Urea treated with the nitrification inhibitor DCD had the poorest AFR (Figure 6). Soil type effects on DCD degradation have been identified (Cahalan *et al.*, 2015 and McGeough *et al.*, 2015). McGeough *et al.* (2015) reported that DCD is less effective in soils with high clay and high organic matter contents. The lowest yields noted for urea+DCD in the present study were associated with weather conditions conducive to volatilisation (Harty *et al.*, in press). Tailoring the nitrification inhibitor rate may address the issue of reduced AFR and yield when urea is treated with DCD. The DCD rate used in the present study may have been too high for best agronomic outcomes in these particular systems and climate. The DCD application method used in these experiments is highly targeted, being incorporated in to the fertiliser granule. Di and Cameron (2005) applied 10 kg/ha DCD in a fine suspended particulate spray across the pasture in a single application. In comparison the rate of DCD applied with the fertiliser at 200 kg N/ha/year was 7 kg/ha/year or



**Figure 5:** Average grassland N uptake for CAN, urea, urea+NBPT across N rates. Fertiliser applied in five equal split applications. Data from six site-years, adapted from Forrestal et al. (in review). ns: not significant \* indicates a treatment effect at  $P < 0.05$ .

1.4 kg DCD/40 kg N/ha application. Despite the relatively low DCD loading rate the overall results of these trials suggest a potential for even lower inclusion rates when DCD is applied in a targeted manner with the fertiliser granule because, as shown later, even at the rate tested DCD was highly effective in reducing  $N_2O$  emissions.

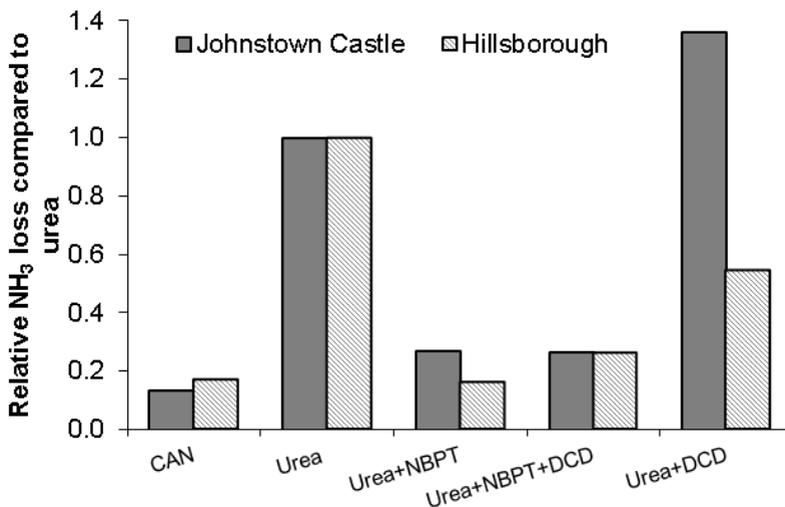


**Figure 6:** Average grassland apparent fertiliser N recovery (AFR). Data from six site-years fertilised at 200 kg N/ha adapted from Harty et al. (in press).

Urea+NBPT could be expected to behave similarly to urea for N loss pathways except for volatilisation. Consequently, the difference in AFR between urea+NBPT and urea gives a crude estimate of apparent  $\text{NH}_3$  loss from urea based on 30 individual applications of fertiliser over six site-years in the temperate maritime climate of Ireland. This loss difference is 4% (Figure 6) and across N rates tested ranges from 4% to 7.6% (Forrestal *et al.*, in review).

### 3.1.3. Ammonia emissions from grassland.

Figure 7 summarises the results of the  $\text{NH}_3$  emissions measurement conducted using wind tunnels. This methodology permits inter-comparison of a large suite of fertilisers in a replicated setting and is useful for determining relative performance and abatement potential of treatments (Forrestal *et al.*, 2016). Ryden and Lockyer (1985) found that wind tunnels can overestimate  $\text{NH}_3$  losses by a factor of 2.4 to 6 during periods of rainfall, hence their usefulness for relative rather than absolute comparison. Our results showed no significant difference in  $\text{NH}_3$  loss from CAN and urea+NBPT in any of ten applications. Both fertilisers had relatively low levels of loss. Ammonia emissions from urea+DCD+NBPT did not differ significantly from CAN in nine out of ten applications, but were significantly higher on one occasion (Forrestal *et al.*, 2016). Compared to untreated urea, the urease inhibitor NBPT reduced  $\text{NH}_3$  loss by 78.5% on average across both sites (Figure 7).



**Figure 7:** Ammonia ( $\text{NH}_3$ ) loss from fertiliser treatments expressed relative to loss from untreated urea. Each histogram indicates the average of five N applications at 40 kg N/ha. Adapted from Forrestal *et al.* (2016).

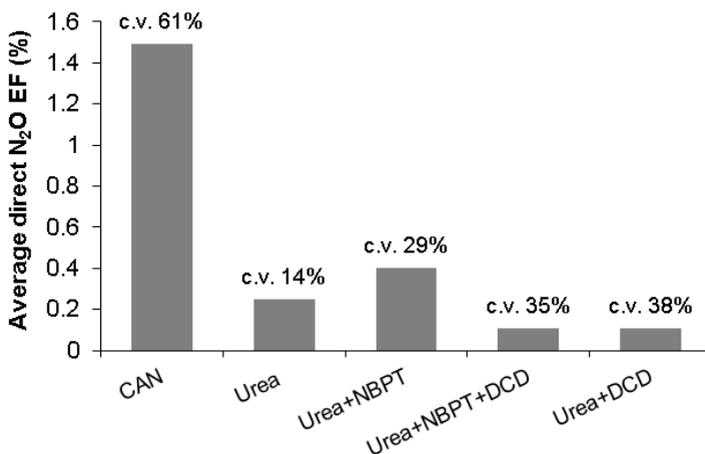
Contrasting effects of DCD on  $\text{NH}_3$  emissions were measured at Johnstown Castle and Hillsborough with greater emissions from urea+DCD compared with urea at Johnstown Castle and lower emissions for urea+DCD at Hillsborough. A meta-analysis by Kim *et al.* (2012) also reported a limited number of cases where DCD decreased  $\text{NH}_3$  emissions but the reduction was much lower than that observed at

Hillsborough. More frequently Kim *et al.* (2012) reported increased or not significantly different  $\text{NH}_3$  emissions due to use of a nitrification inhibitor, as was observed at Johnstown Castle.

Measurements of  $\text{NH}_3$  loss were conducted in spring and summer. It was found that  $\text{NH}_3$  loss from urea can be a significant issue in both spring and summer, indeed the highest loss measured for an individual application occurred in spring (Forrestal *et al.* 2016). This in agreement with Black *et al.* (1985) who noted no marked seasonal pattern of  $\text{NH}_3$  loss from urea in New Zealand grassland. What is clear from our  $\text{NH}_3$  measurements is that the urease inhibitor NBPT is highly effective in reducing  $\text{NH}_3$  emissions from urea to levels comparable to those from CAN. This technology represents a pathway toward reducing  $\text{NH}_3$  losses in agriculture, a key priority of EU national governments committed to reducing national  $\text{NH}_3$  losses.

### 3.1.4. Nitrous oxide emissions from grassland

The Intergovernmental Panel on Climate Change (IPCC) uses the same 1% default  $\text{N}_2\text{O}$  emission factor (EF) irrespective of N fertiliser or soil type. In the present study large differences between fertilisers were observed (Figure 8). Across the six site-years CAN had the highest average direct  $\text{N}_2\text{O}$  EF of 1.49% (Figure 8). Similarly Hyde *et al.* (2016) reported an EF exceeding the IPCC value in Irish grassland. They reported an EF of 2.15% for a single application of CAN. In addition to having the highest EF in the present study; importantly CAN was the most variable treatment, with an across site-year coefficient of variation (c.v.) of 61%. Urea+NBPT and urea had lower direct  $\text{N}_2\text{O}$  emissions and were less variable, with c.v.s of 29% and 14%, respectively (Figure 8).



**Figure 8:** Average  $\text{N}_2\text{O}$  emission factor (EF) and coefficient of variation (c.v.) from six grassland site-years at an annual N fertilisation rate of 200 kg/ha. Adapted from Harty *et al.* (2016).

Thus the urea products reduced the  $\text{N}_2\text{O}$  emission factor and in addition reduced the uncertainty associated with direct EF levels across site-years. The reduction in the

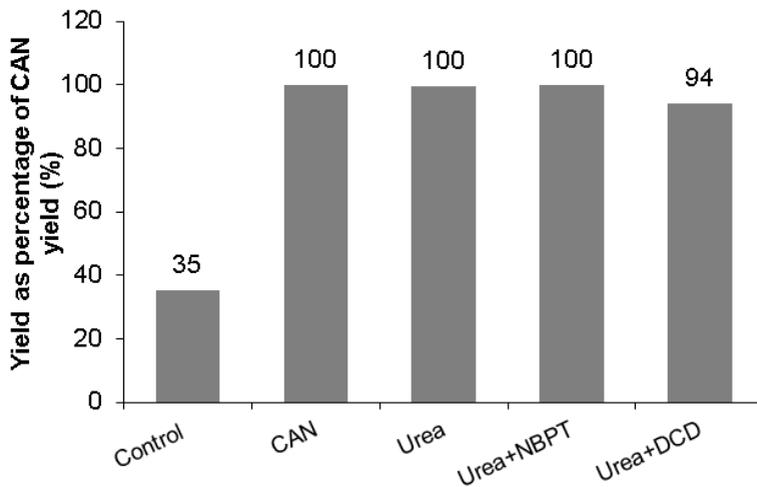
uncertainty is an important point as the precipitation levels and soils of Ireland vary dramatically across the landscape (Figure 1). As a consequence the EF for CAN is likely to be highly variable across sites and seasons. Use of a urea based fertiliser is an option which meets not only the challenge of reducing absolute emissions of N<sub>2</sub>O but also stabilising N<sub>2</sub>O loss outcomes across the landscape in temperate maritime climates such as Ireland.

When the indirect N<sub>2</sub>O emissions associated with urea NH<sub>3</sub> loss were estimated in this study the N<sub>2</sub>O emissions for urea and urea+NBPT were similar (Harty *et al.*, 2016). The results of the present study demonstrate that, compared with CAN, urea amended with the urease inhibitor NBPT is an effective strategy for reducing the N<sub>2</sub>O emissions associated with fertiliser application in temperate maritime Irish grassland. Using urea treated with NBPT the N<sub>2</sub>O reduction can be achieved without causing a large increase in NH<sub>3</sub> emissions, as would be the case if unamended urea usage were increased. Further reductions in N<sub>2</sub>O emissions beyond those possible by substituting urea+NBPT for CAN are technically possible by using a nitrification inhibitor such as DCD. In these experiments the lowest overall N<sub>2</sub>O emissions were for urea treated with both the nitrification inhibitor DCD and the urease inhibitor NBPT, which resolves the NH<sub>3</sub> and associated indirect N<sub>2</sub>O emission difficulty of applying nitrification inhibitors alone.

### 3.2. Spring barley.

#### 3.2.1. Spring barley yields.

Figure 9 summarises the relative yields compared to CAN for three years at a free-draining spring barley site. The data presented is for the 100 kg N/ha rate. This N rate is significantly below the optimum rate for the site and as a consequence crops grown at this N rate are expected to be highly sensitive to differences in crop available N between treatments. Nevertheless, the yields of CAN, urea and urea+NBPT were similar (Figure 9).

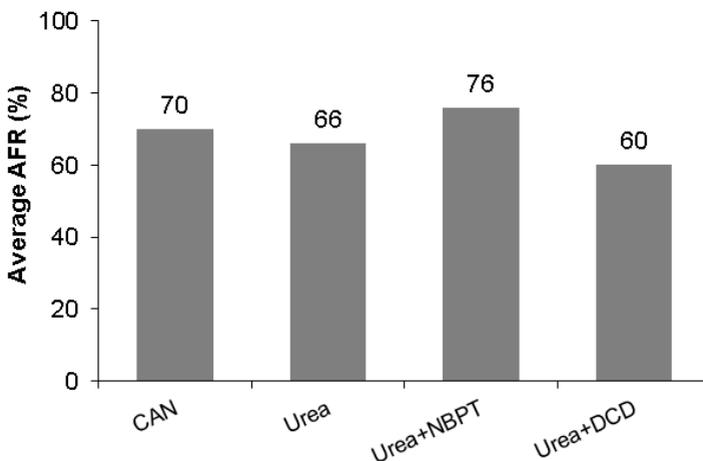


**Figure 9:** *Average spring barley yield expressed as a percentage of CAN yield. Data from three site-years fertilised at 100 kg N/ha. Adapted from Roche et al. (in preparation).*

The finding that urea yielded as well as CAN differs from that of Devine and Holmes (1963) who reported a relative yield of 86% for urea relative to ammonium nitrate for trials conducted on 14 ‘light’ soils. However, they also reported a relative yield of 99% for urea for eight experiments conducted on soils with pH of 7 and below. The soils in the current experiment were also below pH 7. Gately (1994) compared CAN and urea at nine winter wheat sites and showed significant yield reductions with urea at all sites ranging from 0.24 – 0.64 t/ha. Urea with DCD alone had a lower relative yield of 94%, which is consistent with the grassland trials (Figure 4) and evidence in the literature of increased NH<sub>3</sub> loss due to use of nitrification inhibitors (Lam *et al.*, 2016).

### 3.2.2. Spring barley apparent fertiliser N recovery.

The difference in the sensitivity of total N recovery versus yield for detecting differences in the efficiency of different fertilisers can be seen by comparing Figure 9 and Figure 10. In terms of yield CAN, urea and urea+NBPT were similar. However, apparent fertiliser recovery was lowest for urea, which is likely to be a reflection of its potential for NH<sub>3</sub> loss. Urea+NBPT had the highest AFR, although at the 150 kg N/ha rate Roche (in preparation) found no significant difference in N uptake between CAN and urea+NBPT. Urea had lower N uptake than both CAN and urea+NBPT. Based on the 100 kg N/ha rate data presented here, a crude estimate of NH<sub>3</sub> loss can be made by the difference in N recovery of urea compared with CAN and urea+NBPT. This difference was 4-10% on average for these three site-years; very similar to that in the grassland trials.



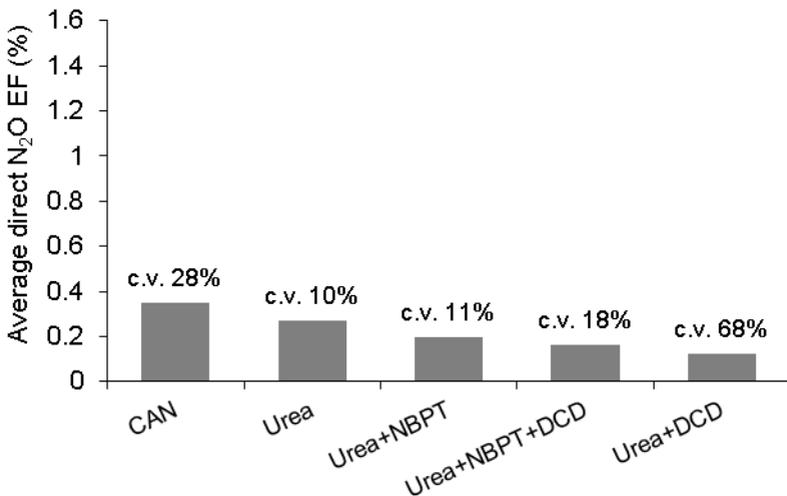
**Figure 10:** Average spring barley apparent fertiliser N recovery (AFR). Data from three site-years at Marshalstown fertilised at 100 kg N/ha.

As was observed in the grassland sites the level of AFR was lowest for urea+DCD. This is counter to findings of increased average yield and nitrogen use efficiency by use of nitrification inhibitors in the meta-analysis of Abalos *et al.* (2014). However, Abalos *et al.* (2014) do state that effectiveness of inhibitors was dependent on the

environmental and management factors of the studies. In the current study, and perhaps surprisingly in the temperate maritime climate of Ireland, the use of DCD alone with urea appears to confer an efficiency disadvantage. As discussed earlier in the grassland section there may be scope to address this efficiency disadvantage by reducing the DCD rate in the granule and by inclusion of NBPT to address the  $\text{NH}_3$  uncertainty associated with adding DCD to urea (Forrestral *et al.* 2016).

### 3.2.3. Nitrous oxide emissions from spring barley

Overall  $\text{N}_2\text{O}$  emissions for all fertilisers tested were relatively low at the Marshalstown spring barley site (Figure 11) in comparison to the grassland sites (Figure 8). The fertiliser  $\text{N}_2\text{O}$  emission factor was less than 0.5% for all fertiliser treatments at this spring barley site. In addition the c.v. for CAN of 28% was lower than the c.v. of 61% observed for the grassland sites. Similar to the grassland sites the use of urea reduced the  $\text{N}_2\text{O}$  c.v. compared to CAN (Figure 11).



**Figure 11:** Average  $\text{N}_2\text{O}$  emission factor (EF) and coefficient of variation (c.v.) over two years at a freely drained spring barley site at an annual N fertilisation rate of 150 kg/ha. Adapted from Roche *et al.* (2016).

The CAN EFs of the present study showing relatively low  $\text{N}_2\text{O}$  emissions from fertiliser N applied to spring barley are consistent with a previous Irish study which reported an EF of 0.5% for CAN during the growing season (Abdalla *et al.*, 2010). There were few differences in  $\text{N}_2\text{O}$  emissions between CAN and urea and urea+NBPT. A comparison between urea and ammonium nitrate (AN) at three UK sites cropped to cereals also found no difference in  $\text{N}_2\text{O}$  emissions between fertiliser CAN and urea (Bell *et al.*, 2015). Similarly, Louro *et al.* (2015) reported no significant effect of N source on  $\text{N}_2\text{O}$  emissions in maize production in the humid Atlantic climate of Galicia, Spain.

The low and less variable EF, compared to the grassland sites, in the present spring barley studies could be explained, at least in part, by the soil characteristics. The soil was a free-draining cambisol with a carbon content of 2.88%, which is typical of Irish soils cropped to spring barley. In a meta-analysis of over 1,000 studies, Stehfest and Bouwman (2006) concluded that N<sub>2</sub>O emissions were significantly lower on soils with soil organic carbon <3% and Gilsanz *et al.* (2016) observed the lowest EFs in soil textures with low clay content (less than 50%) and with sand content greater than 50%.

The inclusion of a nitrification inhibitor has potential to further reduce N<sub>2</sub>O emissions from fertilisers and some reduction was observed in the present study. However, the N<sub>2</sub>O loss reduction potential of nitrification inhibitors is likely to be greatest under high N<sub>2</sub>O loss conditions. The free-draining soils typically cropped to spring barley in Ireland tend to have relatively low levels of N<sub>2</sub>O loss. However, to date little work has been done to examine N<sub>2</sub>O emissions in winter wheat production in the temperate maritime climate of Ireland. The soils cropped to winter wheat generally tend to have poorer drainage characteristics than those soils cropped to spring barley in Ireland. Nevertheless, in the present study the lowest direct N<sub>2</sub>O EF measured was for urea+DCD in 2014 (0%). This outcome indicates potential to grow a spring barley crop fertilised to produce optimal yield but to have the N<sub>2</sub>O emission of an unfertilised crop through the use of fertiliser technologies such as nitrification inhibitors and urease inhibitors to control the NH<sub>3</sub> emission uncertainty of urea.

## 4. CONCLUSIONS.

### 4.1. Yield.

- The yield of CAN and urea was not significantly different in these trials.
- Urea treated with the urease inhibitor NBPT consistently yielded as well as CAN.
- The use of the nitrification inhibitor DCD alone decreased grassland and spring barley yields relative to CAN.
- Addition of NBPT to urea treated with DCD recovered the yield lag caused by the nitrification inhibitor.

### 4.2. Efficiency: apparent fertiliser nitrogen recovery (AFR).

- Urea has the potential for lower (AFR) compared to CAN particularly at higher nitrogen rates.
- Use of the urease inhibitor NBPT ensured that the AFR of urea was consistently at least equal to CAN.
- The nitrification inhibitor DCD used alone had a pronounced negative effect on AFR at the inclusion rate tested in these trials. However, inclusion of NBPT in addition to DCD significantly mitigated this negative effect.

### **4.3. Ammonia.**

- Inclusion of the urease inhibitor NBPT reduced  $\text{NH}_3$  losses from urea by 78.5% on average. As a result  $\text{NH}_3$  loss from urea+NBPT was not significantly different to CAN.
- Variable ammonia loss is a feature of urea usage, however based on comparing the N recovery in plants fertilised with urea, compared to urea+NBPT or CAN,  $\text{NH}_3$  losses are apparently generally low to moderate in temperate Irish grassland and spring barley production.
- Addition of the nitrification inhibitor DCD to urea fertiliser at the rate tested introduces additional uncertainty to the behaviour of urea fertiliser in terms of  $\text{NH}_3$  loss.

### **4.4. Nitrous oxide.**

- Nitrous oxide emissions were highly variable between sites. The free-draining spring barley site had lower emission levels than the grassland sites.
- CAN had the highest direct  $\text{N}_2\text{O}$  emissions in the temperate maritime climate of Ireland, on average exceeding the IPCC default loss. In addition, emissions from CAN were more variable than from urea and urea+NBPT at both the spring barley and grassland sites.
- Urea and urea treated with the urease inhibitor NBPT had lower nitrous oxide emissions than CAN. The magnitude of the loss saving was greatest when emissions were high, with little difference in N forms at the free draining spring barley site but important differences at the grassland sites.
- Addition of the nitrification inhibitor DCD to urea was an effective tool for further suppressing  $\text{N}_2\text{O}$  emissions. At some sites the use of DCD suppressed emissions to levels comparable to the control receiving no N.

The present study found that the fertiliser N form applied along with enhanced efficiency technologies such as urease and nitrification inhibitors are tools which can help to address the key challenge of how to continue to apply fertiliser N to underpin crop yields while curtailing reactive N losses. These trials demonstrate that it is possible to achieve important reductions in nitrous oxide emission, particularly in grassland, without cutting N rates or sacrificing yield or fertiliser efficiency. Options to achieve the  $\text{N}_2\text{O}$  reductions seen in this study by substituting urea+NBPT or urea+NBPT+DCD for CAN in temperate maritime grassland without compromising yield are rare. CAN is generally more expensive than urea as a N source. The resultant price differential provides scope to add urease and/or nitrification inhibitor technologies to urea and remain cost competitive with CAN. As more urease and nitrification inhibitors and formulations enter the market field testing will remain important to evaluate efficacy and to optimise inhibitor rates to meet economic, agronomic and environmental loss mitigation objectives.

## **5. ACKNOWLEDGEMENTS.**

We thank the Department of Agriculture, Food and the Marine (Grant No. 11/S/138), the Agricultural Greenhouse Gas Research Initiative for Ireland (Grant No. 10/RD/SC/716), Department of Agriculture, Environment and Rural Affairs, Northern Ireland and the Walsh Fellowship Scheme for the funding provide to Ms Mary Harty and Ms Leanne Roche. We thank the large team of placement students, technical and farm staff at Teagasc and AFBI along with the farmer James Masterson for access to the spring barley field site.

## 9. REFERENCES

- Abdalla, M., Jones, M., Ambus, P. and Williams, M. (2010). Emissions of nitrous oxide from Irish arable soils: effects of tillage and reduced N input. *Nutrient Cycling Agroecosystems*. **86**, 53-65.
- Abalos, D., Jeffery, S. Sanz-Cobena, A., Guardia, G. and Vallejo, A. (2014). Meta-analysis of the effect of urease and nitrification inhibitors on crop productivity and nitrogen use efficiency. *Agriculture, Ecosystems and Environment*. **189**, 136-144.
- Bell, M., Hinton, N., Cloy, J., Topp, C., Rees, R., Cardenas, L., Scott, T., Webster, C., Ashton, R., Whitmore, A., Williams, J.R., Balshaw, H., Paine, F., Goulding, K.W.T. and Chadwick, D.R. (2015). Nitrous oxide emissions from fertilised UK arable soils: fluxes, emission factors and mitigation. *Agriculture, Ecosystems and Environment*. **212**, 134-147.
- Black, A.S., Sherlock, R.R., Smith, N.P., Cameron, K.C. and Goh, K.M. (1985). Effects of form of nitrogen, season, and urea application rate on ammonia volatilisation from pastures. *New Zealand Journal of Agricultural Research*. **28**, 469-474.
- Cahalan, E., Minet, E., Enfors, M., Mueller, C., Devaney, D., Forrestal, P.J. and Richards, K.G. (2015). The effect of precipitation and application rate on DCD persistence and efficiency in two Irish grassland soils. *Soil Use and Management*. **31**, 367-374.
- Chadwick, D. R., Cardenas, L., Misselbrook, T.H., Smith, K.A., Rees, R.M., Watson, C.J., McGeough, K.L., Williams, J.R., Cloy, J.M., Thorman, R.E. and Dhanoa, M.S. (2014). Optimizing chamber methods for measuring nitrous oxide emissions from plot-based agricultural experiments. *European Journal of Soil Science*. **65**, 295-307.
- Chambers, B. and Dampney, P. (2009). Nitrogen efficiency and ammonia emissions from urea-based and ammonium nitrate fertilisers, *Proceedings International Fertiliser Society*, **657**.
- Chaney, K. and Paulson, G. A. (1988). Field experiments comparing ammonium-nitrate and urea top-dressing for winter cereals and grassland in the UK. *Journal of Agricultural Science*. **110**, 285-299.
- Devine, J.R. and Holmes, M.R.J. (1963). Field experiments on the value of urea as a fertilizer for barley, sugar beet, potatoes, winter wheat and grassland in Great Britain. *Journal of Agricultural Science*. **61**, 391-396.

- Di, H.J. and Cameron, K.C. (2005). Reducing environmental impacts of agriculture by using a fine particle suspension nitrification inhibitor to decrease nitrate leaching from grazed pasture. *Agriculture, Ecosystems and Environment*, **109**, 202-212.
- Dobbie, K.E. and Smith, K.A. (2003). Impact of different forms of N fertilizer on N<sub>2</sub>O emissions from intensive grassland. *Nutrient Cycling in Agroecosystems*. **67**, 37-46.
- Forrestal, P.J., Harty, M., Carolan, R., Lanigan, G.J., Watson, C.J., Laughlin, R.J., McNeill, G., Chambers, B. and Richards, K.G. (2016). Ammonia emissions from urea, stabilised urea and calcium ammonium nitrate: insights into loss abatement in temperate grassland. *Soil Use and Management*. **32**, 92-100.
- Gardiner, M.J., Radford, T., 1980. Soil associations of Ireland and their land use potential. Explanatory bulletin to soil map of Ireland 1980. National Soil Survey of Ireland An Foras Talúntais (The Agricultural Institute), Dublin, Ireland.
- Gately, T.F. (1994). A note on urea versus calcium ammonium nitrate for winter wheat. *Irish Journal of Agricultural and Food Research*. 193-196.
- Gilsanz, C., Baez, D., Misselbrook, T.H., Dhanoa, M.S. and Cardenas, L.M. (2016). Development of emission factors and efficiency of two nitrification inhibitors, DCD and DMPP. *Agriculture, Ecosystems and Environment*. **216**, 1-8.
- Goos, R.J. (2013). A comparison of a maleic-itaconic polymer and N-(n-butyl) thiophosphoric triamide as urease inhibitors. *Soil Science Society of America Journal*. **77**, 1418-1423.
- Harty, M.A., Forrestal, P.J., Watson, C.J., McGeough, K.L., Carolan, R., Elliot, C., Krol, D.J., Laughlin, R.J., Richards, K.G. and Lanigan, G.J. (2016). Reducing nitrous oxide emissions by changing N fertiliser use from calcium ammonium nitrate (CAN) to urea based formulations. *Science of the Total Environment*. 563-564, 576-586.
- Harty, M.A., Forrestal, P.J., Carolan, R., Watson, C.J., Hennessy, D., Lanigan, G.J., Wall, D.P and Richards, K.G. (2016). Temperate grassland yields and nitrogen uptake are influenced by fertilizer nitrogen source. *Agronomy Journal (in press)*.
- Hyde, B.P., Forrestal, P.J., Jahangir, M.M.R., Ryan, M., Fanning, A.F., Carton, O.T., Lanigan, G.J. and Richards, K.G. 2016. The interactive effects of fertiliser nitrogen with dung and urine on nitrous oxide emissions in grassland. *Irish Journal of Agriculture and Food Research*. doi: 10.1515/ijafrr-2016-0001
- IPCC, (2014). Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Core Writing Team, R.K. Pachauri and L.A. Meyer (eds.)]. IPCC, Geneva, Switzerland, 151 pp.
- Irish soil information system. Available online at: <http://gis.teagasc.ie/soils/>
- Keane, G.P., Griffith, J.A. and O'Reilly, J.O. (1974). A comparison of calcium ammonium nitrate, urea and sulphate of ammonia as nitrogen sources for grass. *Irish Journal of Agricultural Science*. **13**, 293-300.
- Kim, D.G., Saggar, S. and Roudier, P. (2012). The effect of nitrification inhibitors on soil ammonia emissions in nitrogen managed soils: a meta-analysis. *Nutrient Cycling in Agroecosystems*. **93**, 51-64.

- Lam, S.K., Sute, H., Mosier, A.R. and Chen, D. Using nitrification inhibitors to mitigate agricultural N<sub>2</sub>O emission: a double edged sword? *Global Change Biology*. doi: 10.1111/gcb.13338.
- Lockyer, D.R. (1984). A system for the measurement in the field of losses of ammonia through volatilization. *Journal of the Science of Food and Agriculture*. **35**, 837-848.
- Louro, A., Báez, D., García, M.I. and Cárdenas, L. (2015). Nitrous oxide emissions from forage maize production on a Humic Cambisol fertilized with mineral fertilizer or slurries in Galicia, Spain. *Geoderma Regional*. **5**, 54-63.
- McGeough, K.L., Watson, C.J., Müller, C., Laughlin, R.J. and Chadwick, D.R. (2016). Evidence that the efficacy of the nitrification inhibitor dicyandiamide (DCD) is affected by soil properties in UK soils. *Soil Biology and Biochemistry*. **94**, 222-232.
- Met Eireann. The climate of Ireland available online at: <http://www.met.ie/climate-ireland/rainfall.asp>
- Meisinger, J.J., Lefcourt, A.M. and Thompson, R.B. (2001). Construction and validation of small mobile wind tunnels for studying ammonia volatilization. *Applied Engineering in Agriculture*. **17**, 375-384.
- Rodgers, G.A., Widdowson, F.V., Penny, A. and Hewitt, M.V. (1984). Comparison of the effects of aqueous and of prilled urea, used alone or with urease or nitrification inhibitors, with those of 'Nitro-Chalk' on ryegrass leys. *Journal of Agricultural Science*. **103**, 671-685.
- Ryden, J.C. and Lockyer, D.R. (1985). Evaluation of a system of wind tunnels for field studies of ammonia loss from grassland volatilisation. *Journal of the Science of Food and Agriculture*. **36**, 781-788.
- Stehfest, E. and Bouwman, L. (2006). N<sub>2</sub>O and NO emissions from agricultural fields and soils under natural vegetation: summarizing available measurement data and modeling of global annual emissions. *Nutrient Cycling in Agroecosystems*. **74**, 207-228.
- Sutton, M.A., Bleeker, A., Howard, C.M., Bekunda, M., Grizzetti, B., de Vries, W., van Grinsven, H.J.M., Abrol, Y.P., Adhya, T.K., Billen, G., Davidson, E.A., Datta, A., Diaz, R., Erisman, J.W., Liu, X.J., Oenema, O., Palm, C., Raghuram, N., Reis, S., Scholz, R.W., Sims, T., Westhoek, H. and Zhang, F.S. (2013). Our Nutrient World: The challenge to produce more food and energy with less pollution. Published by CEH Edinburgh and UNEP Nairobi.
- U.S. EPA, (2012). Global Anthropogenic Non-CO<sub>2</sub> Greenhouse Gas Emissions: 1990-2030 (EPA-R-12-006). Office of Atmospheric Programs, Climate Change Division, U.S. Environmental Protection Agency, 1200 Pennsylvania Avenue, NW, Washington, DC 20460, pp 1-188.
- Watson, C. J. (1987). Factors affecting the efficiency of urea for pot-grown ryegrass. *Journal of Agricultural Science*. **109**, 611-614.
- Watson, C.J., Stevens, R.J. and Laughlin, R.J. (1990). Effectiveness of the urease inhibitor NBPT (N-(n-butyl) thiophosphoric triamide) for improving the efficiency of urea for ryegrass production. *Fertilizer Research*. **24**, 11-15.

- Watson, C.J., Poland, P., Miller, H., Allen, M.B.D., Garrett, M.K. and Christianson, C.B. (1994). Agronomic assessment of  $^{15}\text{N}$  recovery of urea amended with the urease inhibitor nBTPT (N-(n-butyl) thiophosphoric triamide) for temperate grassland. *Plant and Soil*. **161**, 167-177.
- Watson, C.J., Akhonzada, N.A., Hamilton, J.T.G. and Matthews, D.I. (2008). Rate and mode of application of the urease inhibitor N-(n-butyl) thiophosphoric triamide on ammonia volatilization from surface-applied urea. *Soil Use and Management*. **24**, 246-253.

## **RELATED PROCEEDINGS OF THE SOCIETY.**

- 113**, (1970), *Urea-Agronomic Applications*,  
T E Tomlinson..
- 153**, (1976), *Slow Release Fertilisers, Particularly Sulphur-Coated Urea*,  
L H Davies.
- 180**, (1979), *Practical Experience with Ureaform Slow-Release Nitrogen Fertiliser During the past 20 Years and Outlook for the Future*,  
H Schneider, L Veegans.
- 229**, (1984), *Flow of Nitrogen in Grassland*,  
John C Ryden.
- 230**, (1984), *Improving Nitrogen Efficiency in Wetland Rice Soils*,  
Paul L G Vlek, Ian R P Fillery.
- 257**, (1987), *Rationale for Mixed Ammonium Nitrate - Urea Fertilisers and Assessment of Granular Products*, M K Garrett.
- 286**, (1989), *Nitrogen Cycle in UK Arable Agriculture*,  
A E Johnston, D S Jenkinson.
- 287**, (1989), *Nitrogen Transformations and Nitrogen Balances in Scandinavian Soils*, G Bertilsson.
- 298**, (1990), *Ammonia Volatilisation from Agricultural Land*,  
S C Jarvis, B F Pain.
- 299**, (1990), *Losses of Nitrogen by Denitrification and Emissions of Nitrogen Oxides from Soils*, K A Smith, J R M Arah.
- 325**, (1992), *Nitrogen Immobilisation and Leaching in Pasture Soils*,  
S P Cuttle, P C Bourne.
- 326**, (1992), *Nitrogen Economy of Grazed Grassland*,  
M K Garrett, C J Watson, C Jordan, R W J Steen, R V Smith.
- 454**, (2000), *Urease Activity and Inhibition - Principles and Practice*,  
C J Watson.
- 455**, (2000), *Development and Testing of a New Nitrification Inhibitor*,  
W Zerulla, K F Kummer, A Wissemeier, M Rädle.
- 462**, (2001), *Nitrogen Cycling in Grassland Systems*,  
C J Watson.
- 463**, (2001), *Dynamics of Soil and Fertiliser Nitrogen in Arable Systems*,  
S Recous.
- 464**, (2001), *Gaseous Loss of Oxides of Nitrogen from the Agricultural Nitrogen Cycle*, D E Rolston, R T Venterea.
- 573**, (2005), *Foliar Urea Fertilisation and the Management of Yield and Quality in Wheat*, M J Gooding.
- 653**, (2009), *Nitrogen in Physiology - An Agronomic Perspective and Implications for the Use of Different Nitrogen Forms*,  
E A Kirkby, J Le Bot, S Adamowicz, V Roemheld.

- 657**, (2009), *Nitrogen Efficiency and Ammonia Emissions from Urea-based and Ammonium Nitrate Fertilisers*, B J Chambers, P M R Dampney.
- 658**, (2009), *Modification of Nitrogen Fertilisers Using Inhibitors: Opportunities and Potentials for Improving Nitrogen Use Efficiency*, C J Watson, R J Laughlin, K L McGeough.
- 664**, (2009), *Integrating Nitrogen Use and Food Production with Environmental Expectations*, J W Erisman.
- 670**, (2010), *Greenhouse Gas Budgets of Crop Production and the Mitigation Potential of Nutrient Management*, H C Flynn and P Smith.
- 700**, (2011), *Scope for Innovation in Crop Nutrition to Support Potential Crop Yields*, R Sylvester-Bradley and P J A Withers.
- 710**, (2012), *Nitrous Oxide Emissions Associated with Nitrogen Use on Arable Crops in England*, R E Thorman, K E Smith, R M Rees, M Chauhan, G Bennett, S Malkin, D G Munro, R Sylvester-Bradley.
- 728**, (2013), *Comparison of the Environmental Impact of Three Forms of Nitrogen Fertiliser*, S Marquis, T Genter, A Buet, A Berthoud.
- 736**, (2013), *Soil Structure and Greenhouse Gas Emissions*, B Ball, P Hargreaves and J Cloy.
- 753**, (2014), *Spatial Variation of Soil Mineral Nitrogen as a Guide to Sampling for Site-Specific Management*, M A Oliver, C J Dawson.
- 754**, (2014), *Assessment and Prediction of Nitrogen Mineralisation and its Effect on Crop Productivity*, M M A Blake-Kalff, L Blake.
- 760**, (2014), *Comparison of Urea and Ammonium Nitrate in Long-Term Trials: Synthesis of Ten Years of Experimentation*, Ph Eveillard, M Lambert, M Herve, A Bouthier, L Champolivier, S Marquis, C Rocca, D Roussel.
- 762**, (2014), *Ammonia Volatilisation after Application of Fertilisers and Organic Products: Potential for Updating Emission Factors*, S Générmont.
- 773**, (2015), *Nitrogen Use Efficiency (NUE) - An Indicator for the Utilisation of Nitrogen in Agricultural and Food Systems*, O Oenema.

